

# Recent developments in biodegradation of industrial pollutants by white rot fungi and their enzyme system

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**Abstract** Increasing discharge and improper management of liquid and solid industrial wastes have created a great concern among industrialists and the scientific community over their economic treatment and safe disposal. White rot fungi (WRF) are versatile and robust organisms having enormous potential for oxidative bioremediation of a variety of toxic chemical pollutants due to high tolerance to toxic substances in the environment. WRF are capable of mineralizing a wide variety of toxic xenobiotics due to non-specific nature of their extracellular lignin mineralizing enzymes (LMEs). In recent years, a lot of work has been done on the development and optimization of bioremediation processes using WRF, with emphasis on the study of their enzyme systems involved in biodegradation of industrial pollutants. Many new strains have been identified and their LMEs isolated, purified and characterized. In this review, we have tried to cover the latest developments on enzyme systems of WRF, their low molecular mass

mediators and their potential use for bioremediation of industrial pollutants.

**Keywords** Industrial pollutants · Bioremediation · White rot fungi · Lignin mineralizing enzymes · Low molecular mass mediators

## Introduction

In view of the extensive contamination of the environment by persistent and toxic chemical pollutants originating from industrial wastewaters, it is imperative to develop cost effective and efficient methods for their remediation. Bioremediation is a popular and attractive technology that utilizes the metabolic potential of microorganisms to clean up the environment (Watanabe 2001). Recently, the capability of white rot fungi (WRF) for biodegradation of xenobiotics and recalcitrant pollutants has generated a considerable research interest in this area of industrial/environmental microbiology.

WRF are physiological group comprising fungi that are capable of biodegrading lignin and the name white rot derives from the white appearance of the wood attacked by WRF, where lignin removal gives a bleached appearance (Pointing 2001). Taxonomically, WRF are mostly basidiomycetes, although few ascomycetes are also capable of white-rot decay (Eaton and Hale 1993). The extracellular and non-specific lignin

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mineralizing enzymes (LMEs) and low-molecular-mass mediators enhance the bioavailability of pollutants to WRF (Pointing 2001). Property of WRF to withstand a wide range of pH further enhances their pollutant degradation capabilities (Verma and Madamwar 2002).

Many reviews (Kirk and Farrell 1987; Cullen and Kersten 1992; Gold and Alic 1993; Barr and Aust 1994; Hatakka 1994, 2001; Reddy 1995; Pointing 2001; Hofrichter 2002; Martínez 2002; Shah and Nerud 2002; Wesenberg et al. 2003) are available on different aspects of WRF and their enzyme systems. This review will focus on recent developments in biodegradation of industrial pollutants by WRF and their enzyme system.

### Enzyme system of white rot fungi

Lignin peroxidase (LiP), manganese peroxidase (MnP), laccase and versatile peroxidase (VPs) are the major LMEs of WRF involved in lignin and xenobiotic degradation by WRF (Pointing 2001; Wesenberg et al. 2003). Accessory enzymes such as  $H_2O_2$ -forming glyoxal oxidase, aryl alcohol oxidase, oxalate producing oxalate decarboxylase (ODC), NAD-dependent formate dehydrogenase (FDH) and P450 monooxygenase have also been isolated from many WRF strains (Wesenberg et al. 2003, 2005; Doddapaneni et al. 2005; Aguiar et al. 2006).

Lignin peroxidases (LiPs) are capable of mineralizing a variety of recalcitrant aromatic compounds (Shrivastava et al. 2005). The molecular mass of LiPs from different WRF strains varies from 37 to 50 kDa (Asgher et al. 2006; Hirai et al. 2005). pH and temperature activity profiles of LiPs from different sources vary significantly with optimum activities shown between pH 2–5 and 35–55°C, respectively (Yang et al. 2004; Asgher et al. 2007). Immobilization of LiP has been found to enhance its pH and temperature optima as well as thermostability and catalytic properties (Asgher et al. 2007). The natural fungal secondary metabolites veratryl alcohol (VA) and 2-chloro-1,4-dimethoxybenzene act as redox mediators to stimulate the LiP catalyzed oxidation of a wide range of recalcitrant substrates (Teunissen and Field 1998; Christian et al. 2005). Oxidation of VA by YK-LiP2 from *P. chrysosporium* and other LiPs reveal ordered bi–bi ping–pong mechanisms

rather than peroxidase bi–bi ping–pong kinetics shown by Pc-LiP (Liu et al. 2003; Hirai et al. 2005). Kinetic analysis suggests that cationic radical of VA accelerates the catalytic cycle of LiP possibly by conversion of LiP (II) and/or LiP(III) to LiP (Lan et al. 2006). Cationic surfactant cetyltrimethylammonium bromide (CTAB) is an inhibitor for LiP; it lowers the  $V_{max}$  and causes  $K_M$  of  $H_2O_2$  to decrease dramatically and that of VA to increase possibly by modifying the enzyme conformation (Liu et al. 2003). *Phanerochaete chrysosporium* LiP is also inhibited by ethylenediamine tetraacetic acid (EDTA) and *N-N'-N''-N'''*-tetramethylethylenediamine (TEMED). However, in the presence of excess  $Zn^{2+}$  and other metal ions, the EDTA and TEMED mediated non-competitive inhibition of LiP catalyzed veratryl alcohol oxidation can be reversed (Chang and Bumpus 2001). The oxidation by LiP depends on the optimum molar ratio of  $H_2O_2$  to pollutant. At lower concentrations,  $H_2O_2$  is an activator of *Phanerochaete chrysosporium* LiP (Pc-LiP), while at higher concentrations it is an inhibitor and rapidly deactivates the enzyme (Yu et al. 2006; Lan et al. 2006).

Manganese peroxidases (MnPs) are extracellular glycoproteins with an iron protoporphyrin IX (heme) prosthetic group with molecular weights varying between 32 and 62.5 kDa, optimum pH of 4–7 and optimum temperature of 40–60°C (Ürek and Pazarlioglu 2004; Baborová et al. 2006). MnPs are secreted in multiple isoforms in carbon and nitrogen limited media supplemented with  $Mn^{2+}$  and VA (Hakala et al. 2005; Cheng et al. 2007). The marked differences in structure and characteristics (pI 3.4 & 3.9; molecular masses 47 & 52 kDa) and differential regulation of two different MnP encoding genes MnPA and MnPB in *Physiisporinus rivulosus* T24li and their variable properties also confirm the expression of MnP isoforms (Hakala et al. 2005, 2006).

$Mn^{2+}$  performs the role of mediator for MnP. High resolution crystal structure has revealed that MnP catalyzes the peroxide dependent oxidation of  $Mn^{2+}$  to  $Mn^{3+}$  and  $Mn^{3+}$  is released from the enzyme in complex with oxalate or with other chelators (Sundramoorthy et al. 2005). Along with pyrophosphate, the various  $Mn^{2+}$  chelators that enhance the activity are malonate, oxalate, L-tartrate, oxaloacetate, L-malate, and methylmalonate (Mäkelä et al. 2005). MnP can also utilize both  $H_2O_2$  and various organic

hydroperoxides, such as peracetic acid, *m*-chloroperoxybenzoic acid, and *p*-nitroperoxybenzoic acid, as its source of oxidizing equivalents. MnP activity is inhibited by  $\text{NaN}_3$ , ascorbic acid,  $\beta$ -mercaptoethanol and dithreitol (Ürek and Pazarlioglu 2005), whereas the activity can be enhanced in the presence of cooxidants such as glutathione, unsaturated fatty acids and Tween 80 (Hofrichter 2002; Ürek and Pazarlioglu 2005). The stability of MnP can be improved by immobilization with sodium alginate, gelatin or chitosan as carriers and glutaraldehyde as crosslinking agent (Cheng et al. 2007).

Laccases are *N*-glycosylated extracellular blue multicopper oxidases (Wells et al. 2006) with molecular masses varying from 58 to 90 kDa (Murugesan et al. 2006; Salony et al. 2006; Zouari-Mechichi et al. 2006; Quarantino et al. 2007). Purified laccases from *Cyathus bulleri* and *Panus tigrinus* CBS577.79 have shown 16% and 6.9% (high mannose type) glycation, respectively (Salony et al. 2006; Quarantino et al. 2007). The pH and temperature optima of laccases from different WRF vary from 2 to 10 and 40 to 65°C, respectively (Lu et al. 2005; Ullrich et al. 2005; Murugesan et al. 2006; Zouari-Mechichi et al. 2006; D'Souza et al. 2006; Quarantino et al. 2007). Two laccase isozymes LacI and LacII have been identified in *Physisporinus rivulosus* T241i (pI 3.1 & 3.3; MW 66 & 68 kDa), *Trametes trogii* (pI 4.3 & 4.5; MW around 62 kDa; optimum pH 2.0 & 2.5; optimum temperature around 50°C), *Cerrena unicolor* 137 (pI 3.6 & 3.7; MW 64 & 57 kDa; optimum pH 7 & 10; optimum temperature around 60°C) and *Panus tigrinus* (identical optimum pH 7.0 and 60–65°C temperature) (Cadimaliev et al. 2005; Lorenzo et al. 2006; Mäkelä et al. 2006; Michniewicz et al. 2006; Zouari-Mechichi et al. 2006). Two different laccase encoding genes Pr-lac1 and Pr-lac 2 in *Phlebia radiata* 79 are differentially expressed and regulated (Mäkelä et al. 2006). The Pr-lac 2 gene has a very high (66%) sequence similarity with *Trametes versicolor* laccase.

Various fungal metabolites such as *N*-hydroxyacetanilide (NHA), *N*-(4-cyanophenyl)acetohydroxamic acid (NCPA), 3-hydroxyanthranilate, syringaldehyde, ABTS (2,2'-azino-bis(3-ethylbenzthiazoline-6-sulphonate), 2,6-dimethoxyphenol (DMP), violuric acid, 1-hydroxybenzotriazole (HBT), 2,2,6,6-tetramethylpiperidin-*N*-oxide radical, acetovanillone (AV), acetosyringone and acetohydroxamic acid perform the role of mediators of laccase and enhance

pollutant degradation (Geng et al. 2004; Cho et al. 2006; Lu et al. 2007; Minussi et al. 2007; Quarantino et al. 2007). However, *Trametes versicolor* laccase was more effective for denim washing without a mediator as compared to a commercial laccase (Pazarlioglu et al. 2005). Addition of Cu, Cd, Ni, Mo and Mn increase the activity of WRF laccase, in most cases, whereas Ag, Hg, Pb, Zn, sodium azide, sodium chloride and  $\text{H}_2\text{O}_2$  inhibit its activity (Ullrich et al. 2005; Murugesan et al. 2006). However, laccase from *Trametes trogii* is not affected by Cd, Al, Li and Ca (Zouari-Mechichi et al. 2006), whereas the activity of *Pleurotus ostreatus* laccase is enhanced by Pb and Zn (Baldrian et al. 2005). Immobilization of laccase in alginate-chitosan microcapsules increases stability and catalytic activity of the enzyme for repeated use in dye decolorization (Lu et al. 2007).

Versatile peroxidase (VP) is a heme containing structural hybrid between MnPs and LiPs, since they can oxidize not only  $\text{Mn}^{2+}$  but can also oxidize veratryl alcohol, phenolic and nonphenolic and high-molecular weight aromatic compounds including dyes in manganese-independent reactions (Kamitsuji et al. 2004; Pogni et al. 2005; Shrivastava et al. 2005). VPs have been described in species of *Pleurotus* and *Bjerkandera* (Mester and Field 1998; Kamitsuji et al. 2005a, b; Pogni et al. 2005; Honda et al. 2006; Rodakiewicz-Nowak et al. 2006). The deduced amino acid sequence of MnP2 isolated from *P. ostreatus* and VP (VAXPDGVNTA) from a novel strain of *Bjerkandera* sp. (MW 45 kDa) is more than 95% identical to that of a typical VP (VP-PS1) from *P. eryngii* (Moreira et al. 2006). A homologous enzyme, MnP-GY, from recombinant strains *P. ostreatus* ATCC66376 (Kamitsuji et al. 2004) and MnP2 from *P. ostreatus* TM2-10 (Tsukihara et al. 2006) can directly oxidize veratryl alcohol and high-molecular weight compounds, such as Poly R-478 and RNaseA, although much less efficiently than it oxidized  $\text{Mn}^{2+}$  in the absence of a mediator suggesting that the enzyme is a VP having properties of both LiP and MnP.

The molecular architecture of VP from *Bjerkandera adusta* and *Pleurotus eryngii* includes an exposed neutral tryptophan 172 (Trp 172) radical near to the heme prosthetic group, that is responsible for aromatic substrate oxidation and is a putative  $\text{Mn}^{2+}$  oxidation site (Pogni et al. 2007; Tinoco et al. 2007). However, high resolution crystal structure and

side chain orientation studies have shown that Trp164 is the site involved in long-range electron transfer for aromatic substrate oxidation (Pogni et al. 2006). The crystal structures (solved up to 1.3 Å) of *Pleurotus eryngii* VP, before and after exposure to  $Mn^{2+}$ , showed a variable orientation of the Glu36 and Glu40 side chains that, together with Asp175, contribute to  $Mn^{2+}$  coordination (Ruiz-Dueñas et al. 2007). *Bjerkandera adusta* VP requires  $Ca^{2+}$  for long range electron transfer during oxidation of pollutants. Inactivation of VP by  $Ca^{2+}$ -depletion at optimum pH 4.5 due to the differences in the  $Fe^{3+}$  spin states suggested that  $Ca^{2+}$ -depleted VP is able to form the active intermediate compound I but its long range electron transfer is disrupted (Verdín et al. 2006). The decrease in medium polarity by the addition of organic solvents like acetonitrile, dimethylsulfoxide (DMSO), ethanol, and *n*-propanol leads to inhibition of *Bjerkandera fumosa* VP (Rodakiewicz-Nowak et al. 2006). Oxidizing mediators like veratryl alcohol, acetosyringone and 2,2,6,6-tetramethyl-1-piperidinyloxy (TEMPO) enhance the VP catalyzed decolorization of some textile dyes by *Bjerkandera adusta* (Tinoco et al. 2007).

## Bioremediation of industrial pollutants by WRF

### Biodegradation of bleach plant effluents

Bleach plant effluents (BPEs) from the pulp and paper industry generated during bleaching with chlorine-containing chemicals contain toxic chloroorganics and colored compounds that are difficult to decolorize and cause serious environmental problems (D'Souza et al. 2006). Extracellular LMEs from WRF have the potential to degrade highly toxic phenolic compounds from BPEs (Minussi et al. 2007). BPE-decolorizing and dechlorination/detoxification activities of *Phanerochaete chrysosporium*, *Trametes versicolor*, *Fomes lividus* and *Thelephora* sp. and the enzyme systems have been elucidated (Selvam et al. 2002, 2006). It has been demonstrated that laccase and not MnP or LiP plays the major role in decolorizing and detoxifying effluents from pulp and paper industry by WRF (D'Souza et al. 2006; Font et al. 2006). However, some studies indicate that laccase and MnP are the enzymes responsible for BPE decolorization by *Coriolus versicolor*. This

decolorization is directly proportional to the initial color intensities and addition of glucose stimulates color removal (Driessel and Christov 2001).

Immobilization of *Trametes versicolor* in nylon and polyurethane foam cubes causes only 36% reduction in colour, 54% reduction in aromatic compounds and up to 5.7 fold reduction in toxicity as compared to *Trametes versicolor* pellets causing 84.8% colour removal and 70.2% aromatic compounds reduction with laccase as the major enzyme activity (Font et al. 2006). It suggests that immobilization of the fungus causes a significant reduction in laccase production/activity. Acetohydroxamic acid is very efficient mediator for *Trametes versicolor* laccase catalyzed bioremediation of BPEs, reducing 70% and 73% of the total phenol and total organic carbon, respectively as compared to 23% phenol reduction in the presence of HBT or in the absence of any mediator (Minussi et al. 2007).

### Biodegradation/decolorization of dyestuffs

Different dyes and pigments are extensively used in the textile, paper, plastic, cosmetics, pharmaceutical and food industries (Levin et al. 2005). Interest in the biodegradation of synthetic dyes has primarily been prompted by concern over their possible toxicity and carcinogenicity (Maas and Chaudhari 2005; Salony et al. 2006; Revankar and Lele 2007). WRF are better dye-degraders than prokaryotes due to their extracellular non-specific LME system capable of degrading a wide range of dyes (Christian et al. 2005). Most of the earlier dye decolorization studies were based mainly on *Phanerochaete chrysosporium* and *Trametes versicolor* (Toh et al. 2003). However, other WRF including *Phellinus gilvus*, *Pleurotus sajor-caju*, *Pycnoporus sanguineus* (Balan and Monteiro 2001), *Dichomitus squalens*, *Irpex flavus*, *Daedalea flavida*, *Polyporus sanguineus* (Chander et al. 2004; Eichlerová et al. 2006; Chander and Arora 2007), *Funalia trogii* ATCC 200800 (Ozsoy et al. 2005), *Ischnoderma resinosa* (Eichlerová et al. 2006) and *Ganoderma* sp. WR-1 (Revankar and Lele 2007) have been demonstrated to have higher dye decolourization rates than *P. chrysosporium* and *Trametes versicolor*.

LME-producing profiles and patterns of their expression vary among different WRF cultures depending upon the chemical structure and functional groups of the dyes being degraded (Table 1) and this

**Table 1** Lignolytic enzymes of white rot fungi involved in biodegradation of different dyestuffs

WRF strain	Dyestuff	Enzymes	References
<i>Coriolus versicolor</i> f. antarcticus	Malachite Green, Azure B, Poly R-478, Anthraquinone Blue, Congo Red, Xylidine	Laccase	Levin et al. (2004))
<i>Cerrena unicolor</i> (Bull. ex Fr.) Murr. 137	Acid Blue 62, Acid Blue 40, Reactive Blue 81, Direct Black 22, Acid Red 27	Laccase	Michniewicz et al. (2008)
<i>Daedalea quercina</i>	Chicago Sky Blue, Poly B-411, Remazol Brilliant Blue R, Trypan Blue, Reactive Blue 2	Laccase	Baldrian (2004)
<i>Funalia trogii</i>	Reactive BLaack 5	Laccase	Mazmanci and Ünyayar (2005)
<i>Funalia trogii</i>	Remazol Brilliant Blue Royal (RBBR), Drimaren Blue CL-BR	Laccase	Erkurt et al. (2007)
<i>Irpex lacteus</i>	Reactive Blue 19 (RBBR), Reactive Black 5	MnP, laccase	Máximo and Costa-Ferreira (2004)
<i>Irpex lacteus</i>	Reactive Orange 16 and Remazol Brilliant Blue R.	Laccase	Svobodová et al. (2008)
<i>Irpex lacteus</i> (immobilized)	Reactive Orange 16	Laccase, MnP	Tavcar et al. (2006)
<i>Lentinula</i> (Lentinus) <i>edodes</i>	Remazole Brilliant Blue R	MnP	Boer et al. (2004)
<i>Phanerochaete chrysosporium</i> (immobilized on ZrOCl <sub>2</sub> -activated pumice)	Direct Blue 15	MnP	Pazarlioglu et al. (2005)
<i>Phanerochaete chrysosporium</i> (immobilized on Kissiris)	Mythylene Blue	MnP	Karimi et al. (2006)
<i>Phanerochaete chrysosporium</i> BKM-F1767	Direct Blue 15, Direct Green 6, Congo Red	MnP	Ürek and Pazarlioglu (2005)
<i>Pleurotus pulmonarius</i>	Amido Black, Congo Red, Trypan Blue, Methyl Green, Remazol Brilliant Blue R, Methyl Violet, Ethyl Violet, Brilliant Cresyl Blue, Methylene Blue, Poly R-478	Laccase	Tychanowicz et al. (2004)
<i>Pleurotus ostreatus</i>	Phenol Red, Orthocresol Red, Meta-cresol Purple, Bromophenol Red, Bromocresol Purple, Bromophenol Blue, Bromocresol Green	MnP, MiP	Shrivastava et al. (2005)
<i>Scyzyphyllum commune</i> IBL-06	Solar golden yellow R	MnP, laccase	Asgher et al. (2008)
<i>Trametes trogii</i>	Malachite Green, Xylidine, Ponceau 2R, Anthraquinone Blue	Laccase, MnP, GOx	Levin et al. (2005)
<i>Trametes versicolor</i>	Remazol Brilliant Blue R	LiP	Christian et al. (2005)
<i>Trametes versicolor</i> CNPR8107	Remazol Brilliant Blue RR, Remazol Red RR, Remazol Yellow RR	MnP, laccase	Toh et al. (2003)

appears to be of most interest for practical applications (Boer et al. 2004; Kariminiaae-Hamedani et al. 2007). Mechanisms of biodegradation and profiles of biodegradation products of different dyes have also been determined (Zhao et al. 2005, 2006; Gavril and Hodson 2007; Zhao and Hardin 2007).

Decolorization of Direct Blue 15 by *Phanerochaete chrysosporium* immobilized on ZrOCl<sub>2</sub>-activated pumice follows the first-order kinetics with respect to initial dye concentration and MnP plays the major role in decolorization with mycelial adsorption being the minor mechanism (Pazarlioglu et al. 2005). MnP



was also the major enzyme with minor activities of laccase and LiP involved in decolorization of Solar golden yellow R by *Schizophyllum commune*.

Low molecular mass redox mediators like ABTS are necessary for laccase-catalyzed decolorization of most of the dyes (Lu et al. 2005, 2007). However, *Pleurotus pulmonarius* and *Lentinula edodes* SR-1 have been reported to produce only extracellular laccase to decolorize dyes of different spectra without any mediators (Nagai et al. 2002). MnP mediated decolorization of azo dyes Direct Blue 15, Direct Green 6, and Congo red by *Phanerochaete chrysosporium* can be enhanced by the addition of Tween-80 (Ürek and Pazarlioglu 2005) and copper (Tychanowicz et al. 2006). Whereas, LiP produced by *Trametes versicolor* can decolorize Remazol Brilliant Blue R (RBBR) in the presence as well as in the absence of VA (Christian et al. 2005). The decolorization ability of WRF can be substantially increased by carefully optimizing the operational conditions such as initial dye concentration, nutrient content of the media, age of fungus and carbon and nitrogen sources (Ozsoy et al. 2005; Nilsson et al. 2006; Sanghi et al. 2006). Addition of glucose as carbon source caused a dramatic increase in decolorization of Solar golden yellow R by *Schizophyllum commune* IBL-06, whereas additional nitrogen sources were inhibitory to MnP formation and dye decolorization (Asgher et al. 2008).

#### Bioremediation of olive oil mill wastewater

Olive oil extraction produces huge volume of wastewater, known as olive mill wastewater (OMWW) and dry residues of olives (ADOR). Both have high organic load, acidic pH and contain recalcitrant and toxic substances such as phenolic and lipidic compounds that is a serious concern of the olive industry (Ahmadi et al. 2006; Aranda et al. 2006). OMWW has high chemical oxygen demand (COD) of up to 200 g l<sup>-1</sup> and organic fraction includes sugars, tannins, polyphenols, polyalcohols, pectins and lipids. Some of these substances (mainly sugars and polyalcohols) can be used as carbon and energy sources for microbial growth. Conventional biological wastewater treatments are ineffective for OMWW treatment since phenolics possess antimicrobial activity (Ahmadi et al. 2005). Most of the studies have been focused on bioremediation as a means of reducing the polluting effect of OMWW and its bio-transformation into

valuable products (Ramos-Cormenzana et al. 1995). There are reports (Sanjust et al. 1991; Zervakis et al. 1996; Kalmis and Sargin 2004) on the cultivation of *Pleurotus* spp., *Pleurotus sajor-caju* and *P. cornucopiae* var. *citrinopileatus* using wheat straw moistened with mixtures containing 25% and 50% OMWW.

The removal of total phenols from OMWW relative to the total organic load consumed indicates the highest capability for free as well as loofah-immobilized *Phanerochaete chrysosporium* (Garcia et al. 2004; Ahmadi et al. 2005). OMWW can be detoxified through removal of organic matter, decreasing COD/BOD ratio by *Phanerochaete chrysosporium* or *Trametes versicolor* in the presence of complex microbial consortia in combined aerobic/anaerobic systems for its reuse and biogas production on industrial scale (Dhouib et al. 2006). *Pycnoporus coccineus*, *Pleurotus sajor-caju*, *Corioloropsis polyzona* and *Lentinus tigrinus*, are also very active in color and COD removal of OMWW at 50 and 75 g l<sup>-1</sup> COD (Jaouani et al. 2003). At 100 g l<sup>-1</sup> COD only *Pycnoporus coccineus* and *Pleurotus sajor-caju* are effective. *Panus trigrinus* CBS577.79 gives better COD reduction (60.9%), dephenolization (97.2%) and decolorization (75%) of OMWW in bubbled column bioreactor (BCB) as compared to stirred tank reactors (STR) due to possible occurrence of shear stress (D'Annibale et al. 2006).

A better decolourisation of OMWW by *Corioloropsis polyzona* has been reported (Jaouani et al. 2006) under LiP induction conditions (5 mM veratryl alcohol addition) than when LiP was repressed (100 µM Mn<sup>2+</sup> addition). High levels of laccase have a detrimental effect on OMWW decolourisation concomitant to the formation of soluble polymeric aromatic compounds. However, high laccase activity produced by *Pleurotus* spp. in the growth medium reflected a close relationship between the amount of laccase produced and decrease in phenol content (Tsioulpas et al. 2002). *Pycnoporus cinnabarinus* and *Corioloropsis rigida* also produce high laccase activities responsible for 73% phenol reduction of ADOR in 15 days (Aranda et al. 2006).

#### Biodegradation of molasses based wastewater

Fermentation processes using sugarcane molasses yield large volumes of dark brown and highly toxic molasses wastewater (MWW) that contains

considerable amounts of organic compounds. Although most of the organic matter of MWW is removed by means of conventional biodegradation treatments, the removal of dark color due to the presence of melanoidin-type high molecular weight compounds is only marginal (Vahabzadeh et al. 2004). WRF are however, capable of catalyzing degradation of numerous recalcitrant organic compounds often present in MWW (Fu and Viraraghavan 2001; Lacina et al. 2003). The color removal ability of *Phanerochaete chrysosporium* is correlated to the activity of ligninolytic enzymes LiP and MnP (Vahabzadeh et al. 2004). The color reduction in the presence of VA is lower than in its absence, thus confirming the major role of MnP. Increased expression of laccase genes in *Trametes* sp. I-62 (lcc1 & lcc2) and basidiomycetous fungus NIOCC #2a upon exposure to MWW accompanied by enhanced color removal, suggested the involvement of laccase in the melanoidins metabolism (D'Souza et al. 2006; González et al. 2007).

Molasses spent wash (MSW) or digested spent wash (DSW) or alcohol distillery wastewater (WAD) is another wastewater from molasses-based alcohol distilleries (Raghukumar 2002; Chopra et al. 2004). The brown color of MSW is due to the presence of melanoidin pigments, which are highly recalcitrant to biodegradation. These pigments, polycyclic aromatic hydrocarbons (PAHs) like benzo(a)pyrenes and phenols are the causes of its toxicity (Raghukumar et al. 2004; Chairattananakorn et al. 2005). Several species of WRF have been reported to remove about 70–80% of the color present in MSW based effluents (Fahy et al. 1997; Raghukumar 2002; D'Souza et al. 2006). Immobilized mycelia of *Pycnoporus coccineus* on polyurethane foam removed nearly twofold higher WAD color and threefold higher total phenol content than did free mycelia (Chairattananakorn et al. 2005). Decolorization of MSW using free and immobilized mycelia of *Flavodon flavus* is accompanied by simultaneous detoxification and decrease in PAH contents of the MSW possibly via the action of glucose oxidase, accompanied by the production of  $H_2O_2$  that acts as a bleaching agent in the process (Raghukumar and Rivonkar 2001; Raghukumar et al. 2004). Decolorization and COD reduction (52%) of DSW by *Coriolus versicolor* is dependent on the carbon source and addition of organic/inorganic nitrogen has no enhancing effect on decolorization and COD reduction (Chopra et al. 2004).

## Biodegradation of byproducts of rubber industry and waste rubber

The extensive use of rubber products, mainly tyres, and the huge amount of waste rubber material produced as a byproduct is an environmental problem of great concern (Bredberg et al. 2002). Natural rubber waste serum (NRWS) is obtained as a by-product during coagulation of latex and is a major pollutant from the rubber industry (Lau and Subramaniam 1991). Although NRWS is being used as nitrogen and potassium supplement in fertilizers, and as a source of quebrachitol and hevealipin, it is still a menace in rubber-producing countries. Recycling of spent rubber material is problematic due to the vulcanisation, which creates strong sulfur bonds between the rubber molecules (Liu et al. 2000).

Different processes for desulfurization of rubber material and to facilitate the reuse of waste rubber have been developed, including biotechnological processes (Christiansson et al. 1998; Bredberg et al. 2001). Microbial devulcanisation is a promising way to increase the recycling of rubber materials. However several microorganisms tested for devulcanisation are sensitive to rubber additives (Christiansson et al. 2000). However, WRF have the potential to bioremediate NRWS (Atagana et al. 1999). Most of the common rubber additives are aromatic compounds and can be effectively removed by LMEs of WRF. *Resinicium bicolor* is the most effective fungus for detoxification of rubber material, especially the ground waste tire rubber (Bredberg et al. 2002). Treatment of aromatic rubber additives with *Resinicium bicolor* enhances the growth of *Thiobacillus ferrooxidans* bacterium as well as desulfurization compared to the untreated rubber. The co-cultures of aromatic additive degrader WRF with desulfurizing bacteria can thus be an attractive option for developing a potentially viable process for biodegradation of wastes from rubber industry.

## Biodegradation of hormone disrupting compounds

Exposure to alkylsubstituted polynuclear aromatic hydrocarbons (Fragoso et al. 1998), stilbenes (Burnison et al. 1999), genistein (Kiparissis et al. 2003), methoxy-chlor (Lee et al. 2006) and endocrine disrupting chemicals (EDC), nonylphenol (NP) and bisphenol A

(BPA) and the personal care product ingredient triclosan (TCS) (Lee et al. 2005; Cabana et al. 2007) has been associated with a variety of reproductive responses in fish (Kiparisis et al. 2003). Degradation of genistein by *Phanerochaete sordida* YK-624 and detection of the activities of ligninolytic enzymes, MnP and laccase during treatment shows the involvement of WRF extracellular lignolytic system in disappearance of genistein (Tamagawa et al. 2005). MnP, laccase, and the laccase-HBT systems of WRF are also effective in removing the estrogenic activities of bisphenol A (BPA), nonylphenol (NP), 17 $\beta$ -estradiol (E2), ethinylestradiol (EE2) with production of high molecular weight oligomeric metabolites through radical polymerization mechanism and formation of C–C and C–O bonds (Tsutsumi et al. 2001; Suzuki et al. 2003; Lee et al. 2005).

*Stereum hirsutum* has high resistance to methoxychlor and is capable of degrading it by dechlorination and dehydrogenation with a significant reduction in its estrogenic activity (Lee et al. 2006). Removal of NP and BPA is associated with the production of laccase by *T. versicolor* and *Bjerkandera* sp. BOL13 (Soares et al. 2005, 2006). The enhanced biocatalytic elimination of nonylphenol (NP), bisphenol A (BPA) and triclosan (TCS) by *Corioloopsis polyzona* by the addition of ABTS (Cabana et al. 2007) also suggested the involvement of laccase/mediator system.

## Conclusions and future prospects

WRF have tremendous potential for biodegradation of a variety of industrial pollutants. The broad spectrum for biodegradation of pollutants is due to the extracellular and non-specific nature of the enzyme system of WRF, comprising mainly of LiP, MnP, VP, laccase along with other accessory enzymes. The biodegradation capabilities of WRF for different pollutants are variable, mainly due to physiological differences among them, difference in their genetic make up and variable pattern and expression of complex LMEs in the presence of chemically different compounds. The major LMEs from different WRF sources vary widely regarding the number of amino acids, their molecular masses, prosthetic group, low molecular mass mediator requirements, catalytic properties and mechanisms of action on different types of pollutants in performing their role in bioremediation. The

activities of the LMEs can be increased by the addition of different low molecular mass mediators, mostly secreted by WRF themselves. The biodegradation efficiency of WRF can be further enhanced by addition of supplementary nutrients and proper process optimization. Isolation, characterization, immobilization and engineering of LMEs for their hyperactivation and thermostabilization, and their direct use in industrial processes are the area of potential future research.

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